

22 September 2017

Shellharbour City Council PO Box 155 Shellharbour Square Shellharbour City Centre NSW 2529

Attention: **Courtney Williams** Waste Manager Dianne Tierney Waste Manager

Dear Courtney and Dianne,

August 2017 Quarterly Environmental Monitoring – Dunmore Recycling and Waste Disposal Depot, Dunmore, New South Wales.

1.0 Introduction

Environmental monitoring is undertaken on a quarterly basis at the Dunmore Recycling Waste Disposal Depot, Dunmore, NSW (the site), in accordance with Environment Protection Licence (EPL) No. 5984. The monitoring includes sampling groundwater bores, a leachate pond, surface water bodies, a dust gauge and landfill gas at the landfill surface to detect any potential impacts of land filling activities on the environment.

2.0 Scope of works

The quarterly August 2017 monitoring round was undertaken on 21-22 August 2017, (BH14 was sampled on 7 September 2017). During the August 2017 monitoring round groundwater, surface water, leachate, gas and dust samples were collected in and around the site.

Groundwater samples were collected from 10 monitoring bores (BH1b to BH4, BH13, BH14, BH15, BH16, BH19, and BH20). No sample was taken and only standing water level (SWL) was measured at BH9 and BH10. Surface water was collected from the leachate pond (LP1), three on site retention ponds (SWP1, SWP2 and SWP4) and Rocklow Creek (SWC2, SWC_Up and SWC_Down). Sampling locations are shown on Figure 1 (Attachment 1).

A dust gauge bottle was collected to the north of the site (DDG) and a gas walkover of all site buildings and the landfill cap was also undertaken. Landfill gas was measured in the field using an Inspectra Laser Unit (ILU) and a GA5000 Landfill Gas Analyser (GA5000).







3.0 Field measurements

Prior to purging, monitoring bores were measured for SWL. During sampling, field measurements were taken including pH, electrolytic conductivity (EC), oxidation/reduction potential (ORP), dissolved oxygen and temperature. Colour and odour of water samples were also noted. Field measurements recorded for each location are presented in Table 1 (Attachment 2).

All sampling was undertaken in accordance with Environmental Earth Sciences NSW (2011) *Soil, Gas and Groundwater sampling manual.*

4.0 Laboratory analysis

The following analyses were undertaken for site groundwater and surface water during the August 2017 monitoring event:

- groundwater ionic balance (pH, total dissolved salts (TDS), sodium, calcium, potassium, magnesium, fluoride, chloride, ammonium, sulfate, bicarbonate, phosphate and nitrate), total organic carbon (TOC), biological oxygen demand (BOD), total and soluble iron, and soluble manganese;
- surface water (SWC_Up and SWC_Down) ionic balance, total and soluble iron, turbidity, nitrate, ammonium and bicarbonate;
- surface water (SWC2) ammonium, nitrate, bicarbonate and total and soluble iron;
- surface water SWP1, SWP2 and SWP4 ionic balance, total and soluble iron and turbidity;
- additional analyses for SWP4 TOC and BOD; and
- leachate pond ionic balance, TOC, BOD, total and soluble iron, soluble manganese, turbidity, faecal coliforms and E.*Coli*,.

Water samples and the dust sample were sent to Sydney Analytical Laboratories (SAL) for inorganic chemical analyses and to Sonic Healthcare for faecal coliforms and *Escherichia coli* (*E.Coli* counts). Both laboratories are NATA accredited for the methods used.

The inorganic laboratory results for groundwater and surface water are shown in Table 2 and Table 3 (Attachment 2). Calculated ratios of principal ions are presented in Table 4 (Attachment 2).

5.0 Results and discussion

5.1 Groundwater flow

Inferred groundwater contours from the August 2017 standing water level (SWL) measurements are illustrated in Figure 2 (Attachment 1). These were calculated using SWLs from surveyed bores. Groundwater flow direction was towards Rocklow Creek in a southerly direction similar to previous monitoring events (Environmental Earth Sciences NSW, 2010, 2011b, 2012b, 2013, 2014, 2015, 2016 and 2017).

Accumulative rainfall for May 2017 (30.8 mm), June 2017 (42.2 mm) and Jul 2017 (1.4 mm) was 74.4 mm (*BOM – Albion Park Wollongong Airport weather station*). Groundwater levels decreased at all monitoring wells (BH1b, 2, 3, 4, 9, 10, 13, 14, 15 16, 19, and 20). The average of the measured standing water levels throughout the site has decreased by ~0.31 mAHD from 1.23 mAHD in May 2017 to 0.92 m AHD in August 2017. This is thought to be associated with relatively low precipitation rates occurred during the monitoring period.

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5.2 Groundwater

5.2.1 Groundwater sampling locations impacted by leachate

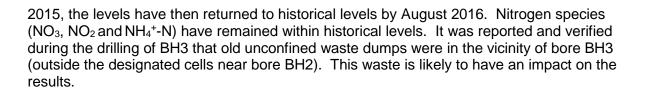
Field and laboratory results from the August 2017 sampling round, specifically from the bores BH1b, BH2, BH3, BH14, BH15, and BH20 displayed chemistry that can be related to leachate impact — BH1b and BH15 showed stronger leachate indicators with high TDS, potassium and ammonium levels. Leachate interaction is demonstrated by elevated concentrations of non-native potassium (K⁺), ammonium (NH₄⁺-N) and nitrate (NO₃⁻) relative to native sodium (Na⁺), calcium (Ca²⁺) and magnesium (Mg²⁺). This comparison is known as the leachate to non-leachate (L/N) ratio.

An L/N ratio >10 may be indicative of leachate impact depending on the combination with other indicators such as odour, colour, BOD and bicarbonate whereas a significant impact is likely at a ratio of >20 (Table 4, Attachment 1).

Bore BH1b is located near the old unlined landfill cell and intercepts leachate within the cell. As such the chemical signature of this bore has historically contained elevated leachate indicators in comparison to other monitoring bores (Schoeller plot BH1 a/b, Attachment 3). This continued during the current monitoring event and the groundwater was found to have an amber colour, and leachate odour noted in combination with elevated TDS (4040 mg/L), K⁺ (220 mg/L) [resulting in low Ca/K – 0.59] and NH₄⁺-N (340 mg/L) concentrations. The absence of oxygen (0.1 ppm Table 1) and presence of soluble Fe²⁺ indicate an anaerobic state and strong biochemical demand in response to microbial respiration. BOD has fluctuated since the bore was installed, ranging from 830 mg/L to 6 mg/L. BODs have remained at similar levels during August 2016, November 2016, February 2017, May 2017, August 2017 rounds and were 8, 7, 7,6 and 6 mg/L respectively. Further evidence of microbial activity / respiration is elevated HCO₃⁻ resulting in a low Cl/ HCO₃⁻ ratio of 0.32 (Table 4). This suggests some degradation of the leachate plume has occurred in this monitoring bore.

Bore BH2 is located down gradient from the old unlined landfill cell. Historically elevated levels of NH_4^+ -N indicate some leachate impact at this location. NH_4^+ -N concentration at BH2 showed an increasing trend since 2010 and reached its historical maximum in August 2017 (49 mg/L). Bicarbonate (HCO₃⁻), Na⁺ and Mg²⁺concentrations in groundwater have shown an increasing trend since January 2008 (Table 2, Schoeller plot BH2, Attachment 3). Chlorine (Cl⁻), Ca²⁺, K⁺ and NH₄⁺-N concentrations slightly increased since the last monitoring round (Table 2, Schoeller plot BH2, Attachment 3) potentially due to an increase in TDS. Low oxygen and negative redox (Table 2) continue to suggest microbial respiration and therefore degradation of the leachate is occurring at this location. Additionally, a leachate odour was noted.

Groundwater from bore BH3 reported slight increases in three native ions (Na⁺, Cl⁻ and Mg²⁺) as well as one non-native ion (K⁺) since last monitoring round. In comparison, NH₄-N and NO₃⁻ reported slight decreases in concentration. The L/N ratio (41.74%) has decreased since the February 2017 monitoring round (60.63%), however due to the low TDS (<1000 mg/L) the L/N ratio must be used with some caution. Long term trends (since 1992) show that K⁺ concentrations generally had an increasing trend from February 2011 until November



ENVIRONMENTAL EARTH SCIENCES

The L/N ratio at bore BH14 showed a decrease in August 2017 round (10.07%) — in May 2017 monitoring round the L/N ratio was 77.58% at this location. Concentrations of K⁺ and NO₃⁻ were lower in the August 2017 monitoring period compared to recent levels (Schoeller plot BH14, Attachment 3). NO₃⁻ concentration (8.7 mg/L) were at its lowest level since February 2012. NH₄⁺-N, on the other hand, was at a peak level (3.4 mg/L) relative to the entire historical data set of this location. Bore BH14 is strategically placed down gradient of landfilling activities and should be continually monitored to determine the water quality in this area.

Bore BH15 displayed an elevated L/N ratio of 87.66% which can be attributed to the elevated K⁺ and NH₄⁺-N concentrations. The K⁺/TDS ratio of 15.70% is high when compared to non-leachate influenced sites generally with K⁺/TDS < 3 (Schoeller plot BH15, Attachment 1). Although elevated the L/N ratio has decreased from the February 2017 monitoring round where the L/N ratio was calculated at 104.84% with a K⁺/TDS ratio of 17.91%. Ammonium (NH₄⁺-N) is elevated at 110 mg/L, compared to other non-impacted locations at the site, which is consistent with previous monitoring rounds. Field observations of a negative redox (negative ORP) and low dissolved oxygen are indicative of a reducing environment. This reducing environment promotes the elevated levels of soluble Mn²⁺ (0.44 mg/L) and Fe²⁺ (14 mg/L). Additionally, bore BH15 is located within a swampy environment where microbiological activity drives reducing reactions that can result in naturally high levels of leachate indicators such as organic carbon and HCO₃⁻. Furthermore, bore BH15 is located near a drainage line with the groundwater bearing zone <0.5 m below the ground surface. Groundwater therefore has the potential to be influenced from local onsite and offsite works and surface water.

Bore BH20 is located down gradient of the landfill, leachate ponds and shallow old landfill. This bore was positioned to assess the chemical characteristics on the boundary of the landfill site. Field observations at bore BH20 recorded a negative redox (-114 mV), clear colour of the groundwater and no odour was detected. The L/N ratio (19.96 %) in August 2017 round remained relatively stable compared the last monitoring round (19.41%). K⁺, on the other hand, showed a slight increase from 26 mg/L (May 2017) to 31 mg/L (August 2017). The TDS remained relatively low (810 mg/L) making the L/N susceptible to natural variations or fluctuations in chemistry. Chemical characteristics of the bore show groundwater is low in Na⁺, with a moderate Ca/K and K/TDS ratio (Table 4). Ammonium levels remain elevated at 15 mg/L however other landfill indicators were low or absent.

5.2.2 Remaining groundwater sampling locations

During the February 2017 monitoring round, ionic chemistry indicated that bores BH4, BH13, BH16 and BH19 only displayed slight to no leachate influence. Chemical composition of each of these bores has been depicted in Schoeller plots in Attachment 3

In the August 2017 monitoring round bore BH4 showed minor leachate influence. Field observations (such as clear water and no odour) and chemical results such as NO_3^- levels below the laboratory limit of detection, relatively low K⁺ and a high Ca/K ratio of 9.06% indicate no or limited leachate influence. Ammonium (NH_4^+-N) levels (6.9 mg/L) remains



above the ANZECC (2000) trigger value for 95% protection of aquatic ecosystems however within the historical range for this bore. The historical chemical composition of the groundwater has remained relatively stable since monitoring began in 1992 (Schoeller plot BH4, Attachment 3). This site is located down gradient of the unlined old landfill cell and will continue to be monitored to assess any potential leachate migration towards Rocklow Creek.

Bore BH13 is located in close proximity to a former night soil area (Figure 1). A slight residual leachate influence has been apparent at this location in the past. Analysis of chemical data from the August 2017 monitoring round shows a decrease of L/N ratio (11.38%) compared to the May 2017 round (16.36%), although the TDS level showed a nearly 10% increase from May 2017 to August 2017. The chemical composition of the groundwater has remained consistent since monitoring began in 2002 (Schoeller plot BH13, Attachment 3).

Bore BH16 is in a swampy area with groundwater field observations recording a Brown/black colour and a sulfuric odour. Recorded ORP indicates a reducing environment (- 77 mV), which may have an influence in the dominance of NH_4^+ -N over NO_3^- . Elevated NH_4^+ -N concentrations (5.9 mg/L) exceeded the ANZECC (2000) criteria, however, the L/N ratio was decreased from 10.25% (May 2017) to 7.92% (August 2017).

Bore BH19 is located on the south west boundary of the site. Field observations included a negative ORP and cloudy white colour. Groundwater chemical characteristics support no/ limited leachate influence with an L/N ratio of 6.34% and a high Ca/K ratio. Bore BH19 is down gradient of current sand dredging activities and unlined landfill cells. Ammonium (NH₄-N) at this location (6.2 mg/L) exceeds the ANZECC (2000) trigger levels. Bore BH19 is positioned to detect any potential leachate migration to the south west of site and will continue to be monitored.

Elevated NH₄-N levels were reported in groundwater across the entire site with bores BH1b (340 mg/L), BH2 (49 mg/L), BH3 (26 mg/L), BH4 (6.9 mg/L), BH13 (2.8 mg/L), BH14 (3.4 mg/L), BH15 (110 mg/L), BH16 (5.9 mg/L), BH19 (6.2 mg/L) and BH20 (15 mg/L) above threshold levels. Nitrate (NO₃⁻) was reported above guidelines (ANZECC 2000) on at BH3 (36 mg/L).

5.3 Surface water monitoring

During Aug 2017 monitoring round samples from Rocklow Creek (SWC2, SWC_Up and SWC_Down) and four surface water ponds (SWP1, SWP2 and SWP4) were collected. Results of surface water analysis (Table 2 and Table 3) indicate that concentrations of ions were within the historical range.

As these ponds are intended to retain any surface water migrating towards Rocklow Creek, the detection of chemical constituents that may be associated with landfill leachate are expected.

The TDS level detected at SWP1 was 1050 mg/L and the nitrogen species were low (NH₄-N at 0.1 mg/L and NO₃⁻ at 0.1 mg/L), which indicate little to no leachate influence.

Surface water sample SWP2 showed minor leachate impact⁻ The surface water pond collects runoff from around the site and potential impacts from site activities are often observed. Ammonium concentration reduced to 2.7 mg/L during this sampling round compared to last May round (7.1 mg/L). Nitrate (NO₃⁻) was below the ANZECC 2000 trigger value with a concentration of 0.53 mg/L. Fluctuating nitrate is common at this location with



previous monitoring events fluctuating between 0.01 and 30 mg/L. All chemical parameters at this location are within historical ranges.

SWP4 displayed ammonium levels of 2.1 mg/L which is above the defined trigger value. Nitrate (NO_3^{-}) levels increased to 18 mg/L exceeding the ANZECC 2000 this is a natural occurrence (Nitrification) and indicates natural attenuation of NH₄-N. All chemical parameters at this location are within historical ranges. Both SWP4 and SWP2 had a decrease in NH₄-N although it should continue to be monitored for any fluctuations in chemical composition indicating a more prominent leachate impact.

The two surface water creek sites SWC_Up and SWC_Down were also sampled during the August 2017 sampling event. These two sites are up- and down-gradient of the established SWC2 site and help assess leachate impacts within Rocklow Creek. SWC_Up and SWC_Down had high concentrations of TDS, notably Na⁺ and Cl⁻ (Table 2), this is due to the tidal nature of these waters and differentiates them from landfill groundwater / surface water.

The low nutrient and L/N ratios within Rocklow Creek indicated that there was no leachate impact within Rocklow Creek. Both SWC_Up and SWC_Down had low concentrations of NH₄⁺-N and NO₃⁻ (>0.1 mg/L and >0.5 mg/L respectively) which did not exceed the ANZECC (2000) trigger value for 95% protection of freshwater ecosystems. Levels detected at all Rocklow Creek locations, are within historical levels established since monitoring was commenced.

5.4 Leachate Pond monitoring

The chemistry of leachate water at the Dunmore Recycling and Waste Disposal Depot is significantly different when compared to the surface and groundwater chemistry of non-leachate influenced bores. This is demonstrated through comparison of chemical data for groundwater from the various bores and the leachate pond presented in Table 2. In particular TDS, TOC, NH₄⁺-N, K⁺, and PO₄ concentrations are generally elevated in leachate pond water compared to other monitoring bores (Schoeller plot Leachate, Attachment 3). Ionic ratios (Table 4) such as low Ca/K (0.40) and high Na/Ca and L/N ratios represent landfill leachate chemical characteristics. These chemical characteristics have been relatively stable over the past 10 years of monitoring.

Laboratory analysis detected *faecal coliforms* and *E.Coli* during this round with concentrations of 400 and 400 CFU/100 ml respectively which is a significant decrease on the previous monitoring round (Table 3). Nevertheless, dermal contact with these waters should be avoided due to health concerns relating to microbial counts.

5.5 Quality assurance/quality control

For quality assurance and quality control the following precision and reliability measures were calculated. The charge balance difference between the summed total of anions against cations (milli-equivalent units) was in the range of 0.45% to 3.14%. The results are a good indication that all major cations and anions present in the groundwater have been analysed and accounted for, providing confidence in the laboratory results obtained.

Field and laboratory practices were further evaluated by comparing the difference between field and laboratory pH and field measured electrical conductivity (EC) against laboratory total dissolved salts (TDS). The range of most relative percent difference (RPD) of field to laboratory pH measurements was between 0.77% and 5.28% (RPD < 10%). The relationship between the field determined EC and laboratory measured TDS relationship

ranged between 0.6 and 0.94. The majority of data is within the TDS/EC typical range of 0.5 to 0.9 and is consistent with historical levels on site (except for SWC_Up and BH19 which yielded TDS/EC ratios of 1.03 and 2.18 respectively).

ENVIRONMENTAL EARTH SCIENCES

RPDs between the intra-laboratory duplicate and the primary sample taken at bore BH20 were all within the acceptable RPD criteria. Thus, the data is considered reliable (Table 6).

5.6 Gas monitoring

Landfill gas was measured in the field using an Inspectra Laser Unit (ILU) and a GA5000 Landfill Gas Monitor (GA5000). Measurements were taken within and around all buildings in a 250 m radius from the current landfill cell as well as across the landfill cap (gas walkover grids of May 2017 round are presented in Figure 3).

All readings were below the site specific criteria outlined in the EPL as the NSW EPA (1996) reporting threshold of 1.25 % v/v CH₄ within onsite buildings and therefore pose no direct risk. Readings were below the threshold concentration for closer investigation and potential action (500 ppm or 0.05 % v/v, NSW EPA [1996], Table 5). No landfill gas was detected with the GA5000. Continued monitoring with both the GA5000 and ILU will be undertaken at quarterly monitoring events.

5.7 Dust

Dust deposition levels to the north of the site were $0.9 \text{ g/m}^2/\text{month}$ total solids, which is below the accepted level of 4 g/m²/month (Australian Standards AS3580.10.1 and AS2724.1). Dust deposition levels to the north of site are within historical ranges and will continue to be monitored to assess the closest sensitive receptor, houses located to the north-west of site.



6.0 Conclusion and recommendations

Groundwater behaviour across the site since the commencement of quarterly monitoring in 1992 has been generally consistent. As the plume beneath the site is generally stable, changes in leachate behaviour into the future are not expected to be significant. Changes to site conditions such as stockpile locations, new landfill cells, new retention ponds and other earth works could potentially impact leachate behaviour on site.

Assessment of monitoring bore BH20 has detected the presence of leachate indicators despite the nearby sampling locations BH19, SWC-UP and SWC-DOWN (Rocklow Creek), which displayed results indicating background characteristics. Although the historical data set of these new locations are relatively limited, it can be said it is likely that on site activities are not significantly impacting Rocklow Creek.

Leachate influenced groundwater may potentially be influencing water quality at bores BH15 and BH16. It is important to note that bore BH15 and BH16 are located in/or near swampy environments or near heavily vegetated areas. Heavily wooded areas to the south can also have a natural attenuation effect on leachate impacted water. Natural attenuation and lower hydraulic gradients in the downstream of BH15 and BH16 are expected to inhibit its rate of migration and should continue to limit its extent of impact on Rocklow Creek. In order to monitor any potential migration from this area, a new surface water sampling location is recommended to be established further downstream in Rocklow Creek (Environmental Earth Sciences, 2017).

Surface water monitoring indicated that on site activities have had limited impact on water quality at locations SWP1, SWP2 and SWP4. Assessment of Rocklow Creek sampling locations (SWC-Up and SWC-Down) reported no concentrations of NH₄-N above the laboratory limit of reporting. These values did not exceed the ANZECC (2000) trigger value for 95% protection of freshwater ecosystems.

Gas concentrations detected at all buildings assessed on site were below guidelines and no action was required to be taken. Gas concentrations on the landfill cap were also within the guidelines. It is recommended monitoring with an ILU and GA5000 continue to the buildings with special attention to the landfill cap due to last May round exceedances.

Depositional dust monitoring results continued to be below guidelines (Australian Standards AS3580.10.1 and AS2724.1) and will continue to be monitored to assess the impact if any dust poses on nearby residential areas.



7.0 Limitations

This letter report has been prepared by Environmental Earth Sciences NSW ABN 109 404 006 in response to and subject to the following limitations:

- 1. The specific instructions received from Shellharbour City Council;
- 2. The specific scope of works set out in PO109055 issued by Environmental Earth Sciences NSW for and on behalf of Shellharbour City Council;
- 3. May not be relied upon by any third party not named in this report for any purpose except with the prior written consent of Environmental Earth Sciences NSW (which consent may or may not be given at the discretion of Environmental Earth Sciences NSW);
- 4. This report comprises the formal report, documentation sections, tables, figures and appendices as referred to in the index to this report and must not be released to any third party or copied in part without all the material included in this report for any reason;
- The report only relates to the site referred to in the scope of works being located at Dunmore Recycling and Waste Disposal Depot located at Buckleys Rd Dunmore, NSW ("the site");
- 6. The report relates to the site as at the date of the report as conditions may change thereafter due to natural processes and/or site activities;
- 7. No warranty or guarantee is made in regard to any other use than as specified in the scope of works and only applies to the depth tested and reported in this report,
- 8. Fill, soil, groundwater and rock to the depth tested on the site may be fit for the use specified in this report. Unless it is expressly stated in this report, the fill, soil and/or rock may not be suitable for classification as clean fill if deposited off site;
- 9. This report is not a geotechnical or planning report suitable for planning or zoning purposes; and
- 10. Our General Limitations set out at the back of the body of this report.

Should you have any further queries, please contact us on (02) 9922 1777.

On behalf of Environmental Earth Sciences NSW

Author Aline Cardoso Environmental Scientist

Project Manager Mert Berberoglu Environmental Scientist

Internal Reviewer Mert Berberoglu Environmental Scientist 112096_Aug_2017_V1



8.0 References

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- NSW Environment Protection Authority (1996). *Environmental guidelines: Solid Waste Landfills.*
- NSW Environment Protection Authority (2016). *Environmental guidelines: Solid Waste Landfills.*

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Scope of services

The work presented in this report is Environmental Earth Sciences response to the specific scope of works requested by, planned with and approved by the client. It cannot be relied on by any other third party for any purpose except with our prior written consent. Client may distribute this report to other parties and in doing so warrants that the report is suitable for the purpose it was intended for. However, any party wishing to rely on this report should contact us to determine the suitability of this report for their specific purpose.

Data should not be separated from the report

A report is provided inclusive of all documentation sections, limitations, tables, figures and appendices and should not be provided or copied in part without all supporting documentation for any reason, because misinterpretation may occur.

Subsurface conditions change

Understanding an environmental study will reduce exposure to the risk of the presence of contaminated soil and or groundwater. However, contaminants may be present in areas that were not investigated, or may migrate to other areas. Analysis cannot cover every type of contaminant that could possibly be present. When combined with field observations, field measurements and professional judgement, this approach increases the probability of identifying contaminated soil and or groundwater. Under no circumstances can it be considered that these findings represent the actual condition of the site at all points.

Environmental studies identify actual sub-surface conditions only at those points where samples are taken, when they are taken. Actual conditions between sampling locations differ from those inferred because no professional, no matter how qualified, and no sub-surface exploration program, no matter how comprehensive, can reveal what is hidden below the ground surface. The actual interface between materials may be far more gradual or abrupt than an assessment indicates. Actual conditions in areas not sampled may differ from that predicted. Nothing can be done to prevent the unanticipated. However, steps can be taken to help minimize the impact. For this reason, site owners should retain our services.

Problems with interpretation by others

Advice and interpretation is provided on the basis that subsequent work will be undertaken by Environmental Earth Sciences NSW. This will identify variances, maintain consistency in how data is interpreted, conduct additional tests that may be necessary and recommend solutions to problems encountered on site. Other parties may misinterpret our work and we cannot be responsible for how the information in this report is used. If further data is collected or comes to light we reserve the right to alter their conclusions.

Obtain regulatory approval

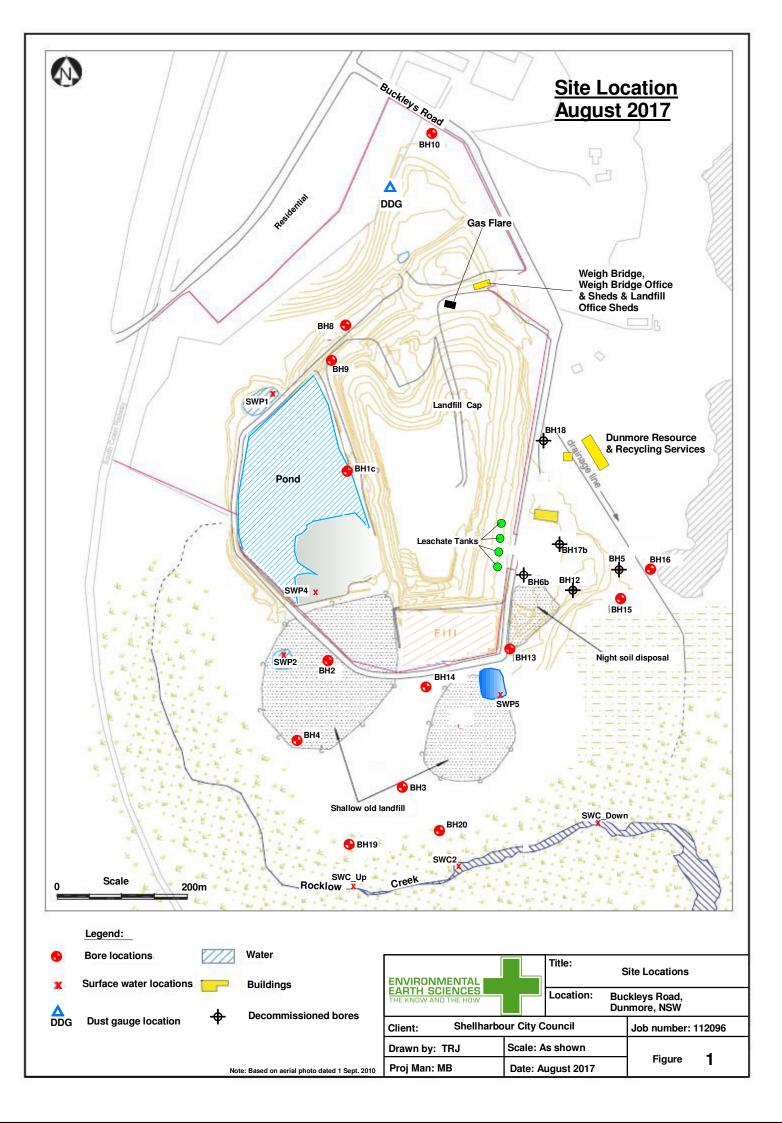
The investigation and remediation of contaminated sites is a field in which legislation and interpretation of legislation is changing rapidly. Our interpretation of the investigation findings should not be taken to be that of any other party. When approval from a statutory authority is required for a project, that approval should be directly sought by the client.

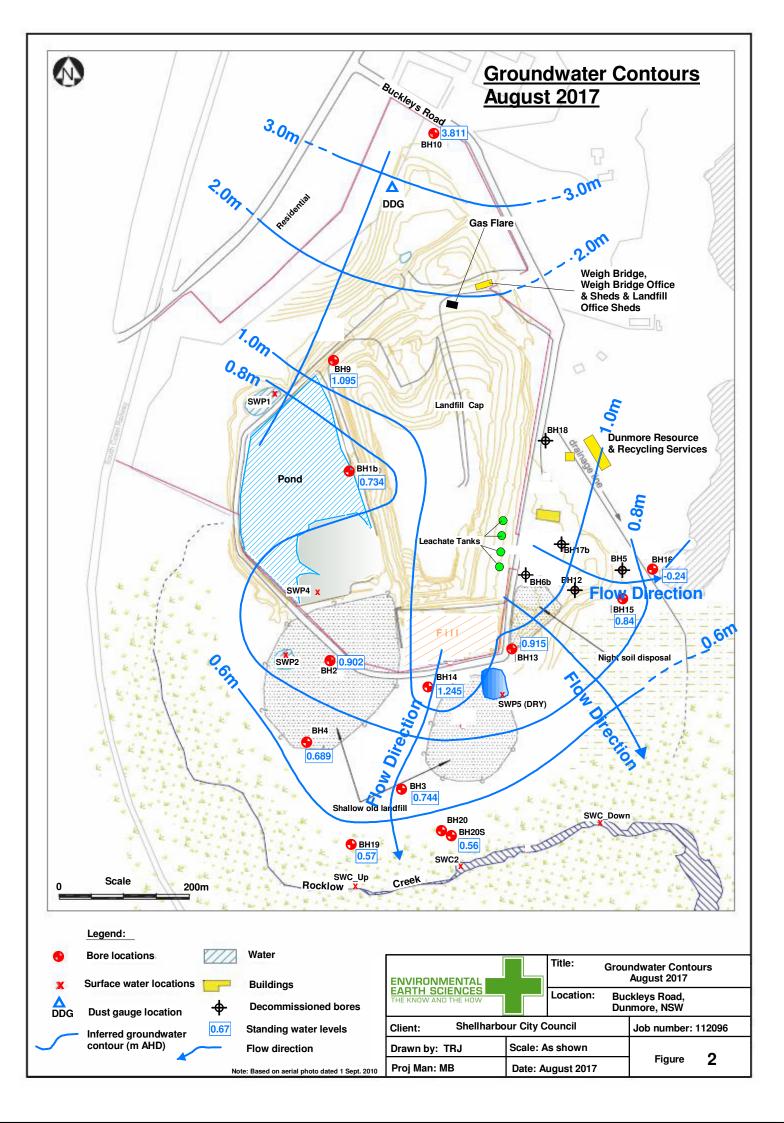
Limit of liability

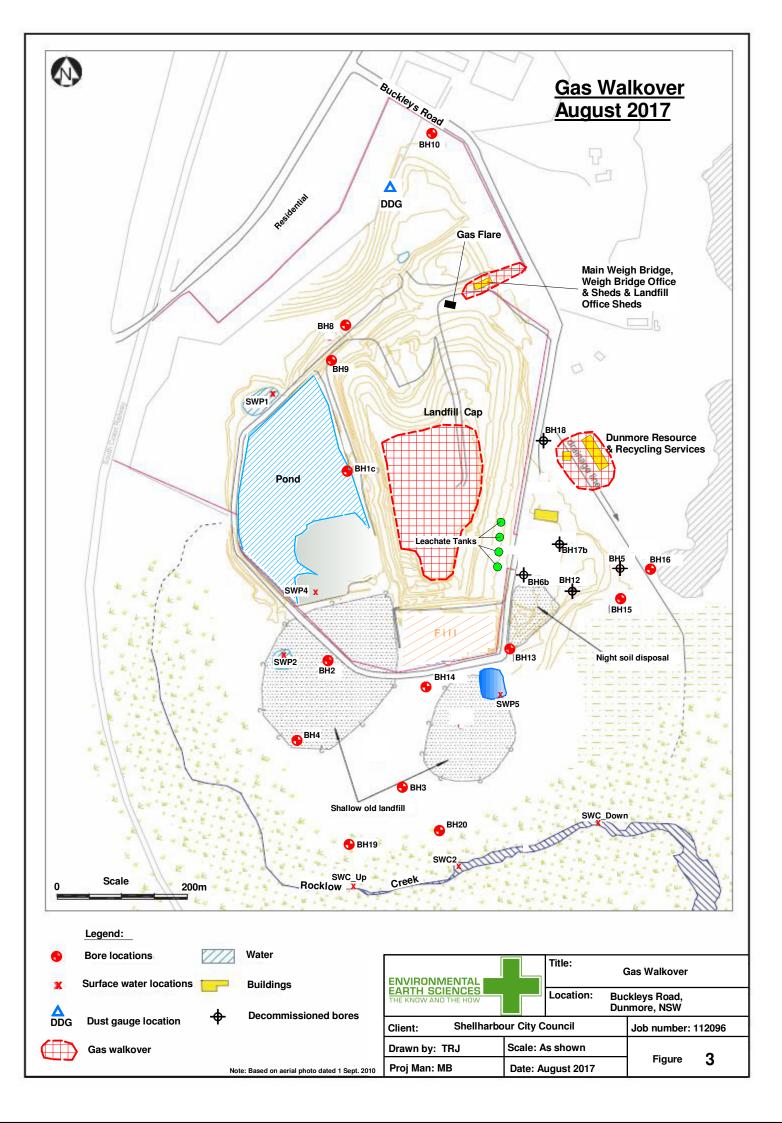
This study has been carried out to a particular scope of works at a specified site and should not be used for any other purpose. This report is provided on the condition that Environmental Earth Sciences NSW disclaims all liability to any person or entity other than the client in respect of anything done or omitted to be done and of the consequence of anything done or omitted to be done by any such person in reliance, whether in whole or in part, on the contents of this report. Furthermore, Environmental Earth Sciences NSW disclaims all liability in respect of anything done or omitted to be done and of the consequence of anything done or omitted to be done and of the consequence of anything done or omitted to be done and of the consequence of anything done or omitted to be done by the client, or any such person in reliance, whether in whole or any part of the contents of this report of all matters not stated in the brief outlined in Environmental Earth Sciences NSW's proposal number and according to Environmental Earth Sciences general terms and conditions and special terms and conditions for contaminated sites.

To the maximum extent permitted by law, we exclude all liability of whatever nature, whether in contract, tort or otherwise, for the acts, omissions or default, whether negligent or otherwise for any loss or damage whatsoever that may arise in any way in connection with the supply of services. Under circumstances where liability cannot be excluded, such liability is limited to the value of the purchased service.

ATTACHMENT 1 FIGURES







ATTACHMENT 2 TABLES

TABLE 1 FIELD MEASUREMENTS – AUGUST 2017

Sample	SWL	SWL	рН	EC	ORP	Temp.	DO	Colour	Odour
Units	mAHD	Dip (m)	-	mS/cm	mV	℃	ppm	-	-
BH1b	0.734	3.22	6.79	7.354	-128	26.2	0.1	Amber	Leachate
BH2	0.902	3.89	6.62	3.327	-85	22.2	0.33	Slight yellow	Leachate
BH3	0.744	3.02	6.96	1.286	-94.1	18.9	2.82	clear	no odour
BH4	0.689	4.33	6.90	1.292	-74.6	19	0.72	clear	no odour
BH9	1.095	3.29	-	-	-	-	-	-	-
BH10	3.811	0.98	-	-	-	-	-	-	-
BH13	0.915	4.38	6.64	1.597	-56	21	-	clear	no odour
BH14	0.465	5.25	-	1.4	-	21	-	clear	no odour
BH15	0.84	0.57	6.69	5.5	-104	13.9	0.87	brown	Leachate
BH16	-0.24	1.62	6.38	2.742	-77	14.3	0.43	brown/black	H2S
BH19	0.57	4.62	6.82	0.4847	-38.1	18.6	0.5	light grey	no odour
BH20	0.56	2.21	7.07	1.148	-110	18.7	1.87	clear	no odour
LP1	-	-	7.61	13.171	-114	18.9	5.97	amber	Leachate
SWC2	-	-	6.97	37.38	87.4	13.8	7.52	clear	no odour
SWC-Up	-	-	7.33	31.573	30.7	12.9	5.77	clear	no odour
SWC-Down	-	-	7.25	38.499	8.7	13.3	4.48	clear	no odour
SWP1	-	-	7.57	1.323	30.2	11	7.39	clear	no odour
SWP2	-	-	7.49	2.462	30.2	12.5	7.81	clear	no odour
SWP4	-	-	7.86	1.655	9.7	13.4	7.43	clear	no odour
SWP5	-	-	-	-	-	-	-	-	-

Notes:

SWL Standing Water Level, measured to the top of the monument or casing; RL – reference level;
 not measured;
 N/A = Not applicable
 DO = dissolved oxygen;
 ORP = electron activity; and
 EC= electrolytic conductivity

TABLE 2 WATER LABORATORY RESULTS – AUGUST 2017

Sample	рН	TDS mg/L	Na mg/L	Ca mg/L	K mg/L	Mg mg/L	NH₄-N mg/L	Cl mg/L	F mg/L	NO₃ mg/L	NO₂ mg/L	SO₄ mg/L	HCO₃ mg/L	PO₄ mg/L	TOC mg/L	BOD mg/L	Sol. Mn mg/L	Sol. Fe mg/L	Tot. Fe mg/L
BH1b	7	4040	705	130	220	120	340	955	0.19	<0.1	-	15	2990	0.49	190	6	0.11	2.2	16
BH2	6.9	2000	340	220	50	79	49	505	0.2	0.13	-	91	1310	<0.1	67	<2	0.43	0.96	13
BH3	7.2	830	64	140	34	26	26	190	0.12	36	-	93	420	<0.1	15	2	0.16	0.03	3.5
BH4	7.1	810	105	145	16	30	6.9	145	<0.1	0.27	0.16	84	550	<0.1	15	3	0.15	0.36	4.3
BH5	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-
BH10	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-
BH13	7	1110	100	210	35	45	2.8	115	0.25	2.6		220	710	<0.1	23	2	0.24	0.12	3.6
BH14	6.8	1260	155	205	29	48	3.4	200	0.55	8.7	<0.1	175	790	<0.1	27	4	0.2	0.12	0.73
BH15	6.6	3950	575	170	620	88	110	1480	0.16	<0.1	<0.1	190	1180	1.2	215	2	0.44	14	22
BH16	6.6	2130	420	200	50	87	5.9	625	0.14	<0.1	-	250	860	0.12	59	3	0.12	0.38	1.9
BH19	7.1	1060	90	205	15	41	6.2	190	<0.1	<0.1	-	140	625	<0.1	15	3	0.08	0.11	41
BH20	7.4	810	56	140	31	35	15	125	0.12	<0.1	-	160	470	0.34	19	2	0.07	0.06	1.8
LP1	7.7	8890	1450	195	485	95	1080	1740	0.47	<0.1	-	120	7510	22	700	75	0.46	2.9	3.3
SWC2	-	-	-	-	-	-	0.1	-	-	0.27	0.2	-	190	-	-	-	-	0.1	1.5
SWP1	7.4	1050	240	74	30	45	<0.1	290	0.22	<0.1	-	135	500	0.12	-	-	-	0.11	0.26
SWP2	7.6	1500	310	135	35	61	2.7	365	0.19	0.53	-	240	715	0.18	-	-	-	0.06	1.1
SWP4	7.8	1300	285	81	30	54	2.1	345	0.31	18	-	230	455	<0.1	31	<2		0.32	0.33
SWP5	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-
SWC-UP	7.2	32600	9620	430	480	1200	0.1	18000	0.42	0.27	-	2260	150	<0.1	-	-	-	0.16	0.78
SWC-DOWN	7.4	33600	9890	410	465	1170	0.1	18500	0.44	0.22	-	2350	225	<0.1	-	-	-	0.13	0.4

Notes:

results and guidelines are expressed in mg/L
 SWC_Do – SWC_Down;
 - not analysed;
 guidelines levels from ANZECC (2000) – Australian and New Zealand guidelines for fresh and marine water quality for the protection of aquatic ecosystems;
 * - guideline from freshwater trigger values as total NH₄-N at different pH values - Table 8.3.7 of ANZECC (2000) - based on average laboratory pH of 7.3 from pH values presented above;
 # - # - based on the recalculated trigger value for freshwater, Hickey 2013; and
 values above the guidelines are **bolded**.



TABLE 3 SURFACE WATER RESULTS - AUGUST 2017

Sample	NH4-N	HCO ₃	Sol. Fe	Tot Fe	FCs	E. Coli
Units	mg/L	mg/L	mg/L	mg/L	CFU/100ml	CFU/100ml
LP1	1080	7510	2.9	3.3	400	400
SWC2	0.1	190	0.1	1.5	-	-
SWC-UP	0.1	150	0.16	0.78	-	-
SWC-Down	0.1	225	0.13	0.4	-	-
SWP1	<0.1	500	0.11	0.26	-	-
SWP2	2.7	715	0.06	1.1	-	-
SWP4	2.1	455	0.32	0.33	-	-
SWP5	_	_	_	-	-	_
ANZECC 2000	1.88*	-	0.3#	-	-	-

Notes:

- = not analysed; 1.

2. FCs = faecal coliforms;

3. E. Coli = Escherichia coli;

guidelines levels from ANZECC (2000) – Australian and New Zealand guidelines for fresh and marine water quality for the protection of aquatic ecosystems; 4.

5. * = guideline from marine trigger values as total NH4-N at different pH values - Table 8.3.7 of ANZECC (2000) - Table 8.3.7 of ANZECC (2000) - based on average laboratory pH of 7.3 from pH values presented in Table 1; # = interim indicative working level presented in section 8.3.7 of ANZECC 2000 (based on Canadian derived

6. guidelines); and

7. values above the guidelines are bolded.

TABLE 4 RATIOS OF PRINCIPAL IONS – AUGUST 2017

							K/TDS	L/N
Bore	Na/CI	Na/Ca	Mg/Ca	Ca/K	CI/SO ₄	CI/HCO₃	(%)	(%)
BH1b	0.74	5.42	0.92	0.59	63.67	0.32	5.45	58.65
BH2	0.67	1.55	0.36	4.40	5.55	0.39	2.50	15.51
BH3	0.34	0.46	0.19	4.12	2.04	0.45	4.10	41.74
BH4	0.72	0.72	0.21	9.06	1.73	0.26	1.98	8.33
BH5	-	-	-	-	-	-	-	-
BH10	-	-	-	-	-	-	-	-
BH13	0.87	0.48	0.21	6.00	0.52	0.16	3.15	11.38
BH14	0.78	0.76	0.23	7.07	1.14	0.25	2.30	10.10
BH15	0.39	3.38	0.52	0.27	7.79	1.25	15.70	87.66
BH16	0.67	2.10	0.44	4.00	2.50	0.73	2.35	7.92
BH19	0.47	0.44	0.20	13.67	1.36	0.30	1.42	6.34
BH20	0.45	0.40	0.25	4.52	0.78	0.27	3.83	19.96
LP1	0.83	7.44	0.49	0.40	14.50	0.23	5.46	89.95
SWC2	-	-	-	-	-	-	-	-
SWP1	0.83	3.24	0.61	2.47	2.15	0.58	2.86	-
SWP2	0.85	2.30	0.45	3.86	1.52	0.51	2.33	-
SWP4	0.83	3.52	0.67	2.70	1.50	0.76	2.31	-
SWP5	-	-	-	-	-	-	-	-
SWC-UP	0.53	22.37	2.79	0.90	7.96	120.00	1.47	-

Notes:

1.

SWC_Do – SWC_Down; % indicates ratios are presented in percentage in that column; and L/N = leachate/non-leachate ratio ; [(K + NH4 + NO3 + NO2)/(Ca + Mg + Na)] x 100. 1. 2. 3.

TABLE 5 SUMMARY OF GAS ANALYSIS – AUGUST 2017

Location	GA 5000 V/V%	ILU V/V%		
Landfill cap	-	0.00063		
Main weigh bridge, weigh bridge office and landfill office sheds	-	0.00026		
Dunmore Resource & Recycling Services	-	0.00036		
GUIDELINES	1.25 % v/v / 0.05 % v/v	1.25 % v/v / 0.05 % v/v		

Notes:

1.

results and guidelines are expressed in V/V %; Guidelines are as per the NSW EPA (1996) reporting accumulation value of 1.25 % v/v CH₄; and surface emission trigger value (500 ppm or 0.05 % v/v); and values above the guidelines are **bolded**. 2.

3.

TABLE 6 QA/QC – AUGUST 2017

Analytes	BH20	SD1	RPD (%)
рН	7.40	7.40	0.00
TDS	795	810	1.87
Na+	52	56	7.41
Ca++	150	140	6.90
K+	32	31	3.17
Mg++	34	35	2.90
CI-	125	125	0.00
F-	0.12	0.12	0.00
NO3-	<0.1	<0.1	-
SO4	160	160	0.00
HCO3-	0.05	0.05	0.00
PO4	0.43	0.34	23.38
NH4-N	15	15	0.00
BOD	1	1	0.00
тос	20	19	5.13
Fe.T	1.7	1.8	5.71
Fe.D	0.05	0.06	18.18

Notes:

results are expressed in mg/L;
 RPD – Relative Percentage Difference
 NA - not analysed;
 values requiring further investigation are **bolded**.

ATTACHMENT 3 SCHOELLER PLOTS

